Distribution and abundance of Chubut Steamerducks: an endemic species to Central Patagonia, Argentina

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Summary

We documented the breeding distribution and estimated abundance of Chubut Steamerducks *Tachyres leucocephalus*, a flightless waterbird endemic to a relatively small section of coastline in Patagonia, Argentina. The distribution of Chubut Steamerducks is restricted to approximately 700 km of coast. We counted 1,703 adult steamerducks at a subset of shorelines within their range and estimated 1,841 adults after correcting for visibility for shore-based surveys. To estimate adult densities in unsurveyed areas, we used two different methods of extrapolation, resulting in estimates of 1,587 and 1,832 adults. Combined with numbers from surveyed shorelines, the total breeding population size is estimated to be between 3,428 and 3,673 adults. In addition, we counted 1,899 juvenile steamerducks, which occur in irregular aggregations. The Interjurisdictional Marine Park in San Jorge Gulf contains about 46% of the entire population, which may provide some protection from disturbance and habitat destruction. However, oil pollution, other anthropogenic activities, and invasive species still pose potential threats to the population.

Resumen

Hemos documentado el rango de distribución y estimado el tamaño poblacional del pato vapor cabeza blanca *Tachyres leucocephalus*, un ave acuática no voladora endémica de una estrecha sección costera de Patagonia, Argentina. La distribución del pato vapor cabeza blanca está restringida a aproximadamente 700 km de costa. Hemos contado 1,703 individuos a lo largo de sectores costeros discretos dentro de su rango de distribución, y estimado un total de 1,841 adultos luego de aplicar los correspondientes factores de corrección. Para estimar la densidad de adultos en los sectores no relevados, hemos utilizado dos métodos diferentes de extrapolación, resultando en un número estimado de 1,587 y 1,832 adultos. De esta forma, al sumar individuos relevados y estimados, el tamaño poblacional oscilaría entre los 3,428 y 3,673 adultos reproductores. Asimismo, hemos contado un total de 1,899 juveniles de pato vapor cabeza blanca, los cuales ocurren en agregaciones regulares. El Parque Interjurisdiccional Marino Golfo San Jorge alberga alrededor del 46% de la población, lo cual podría representar una medida de protección frente a disturbios o destrucción del hábitat. Sin embargo, la contaminación por hidrocarburos y otras actividades de origen antropico junto con la introducción de especies, representan un potencial peligro para la población completa.

Introduction

Estimates of population size are a crucial tool for setting conservation priorities (Karanth and Nichols 1998, Kéry 2002, Thompson 2002). This information allows for the assessment of future changes in population size and hence evaluation of the impacts of habitat loss, pollution, or other
threats (Buckland et al. 2008). In addition, numerical estimates as well as population trends and distribution are important criteria of the IUCN Red List classification scheme (IUCN 2001).

Steamerducks (genus *Tachyeres*) are large diving ducks limited in distribution to southern South America. Four species are currently recognised (Weller 1976, Humphrey and Thompson 1981), including three flightless species: Chubut Steamerduck *Tachyeres leucocephalus*, Magellanic Steamerduck *T. pteneres*, and Falkland Steamerduck *T. brachypterus*, and one flying species: Flying Steamerduck *T. patachonicus*. Flying Steamerducks breed in both freshwater and marine habitats; the three flightless species are strictly marine throughout the annual cycle.

All species of *Tachyeres* are monogamous and form lasting, perhaps lifelong pair bonds (Humphrey and Livezey 1985). Males are extremely territorial and fight frequently (Livezey and Humphrey 1985). All members of the genus feed primarily on large molluscs and crustaceans obtained from the bottom by diving and shallow-water foraging, or found exposed during low tide (Murphy 1936, Weller 1972, Livezey 1989).

Chubut Steamerduck is endemic to the coast of central Patagonia, Argentina (Madge and Burn 1988). There is little information about the distribution or abundance of the species beyond speculative estimates. Numbers have been believed to be fewer than 10,000 individuals (Wetlands International 2002). However, Wetlands International (2002) also states that “the population may only number a few hundred (D. Scott pers. obs.)”. This uncertainty regarding distribution and basic population size hinders adequate design and implementation of conservation and management actions. Chubut Steamerducks are highly territorial, particularly during the breeding season, which is thought to be an important defence of nest site, food, and broods (Vallentin 1924, Weller 1972, 1976). This territorial behaviour may lead to dispersion of nesting pairs and thus may limit local densities. Large flocks of non-breeding birds, primarily juveniles, are excluded from territories and occur in communal aggregations (Murphy 1936, Pettingill 1965, Weller 1972, 1976). We conducted the first surveys of the Chubut Steamerduck to (1) document the breeding range of the species, and (2) estimate abundance, which will allow determination of population trends in the future. These basic attributes constitute critical information for directing subsequent conservation for this poorly-known species.

**Methods**

**Study area and census techniques**

Our study area corresponded to Chubut Province, Patagonia, Argentina, whose coastline is 1,490 km in length. To determine the breeding distribution of Chubut Steamerducks and estimate their abundance, we surveyed 732 km of mainland (divided in segments along the whole Chubut coast) and 103 km of island coast (58 islands and islets; Figure 1). To accurately determine the breeding range, we surveyed coastal sectors located at both northern and southern extremes of the study area to confirm there were no breeding pairs. Surveys were conducted from 2006 to 2008 during September to December, which is the breeding season of Chubut Steamerducks. Although the Flying Steamerduck has been documented in coastal Chubut, we are confident that our observations reflect the abundance and distribution of the much more common Chubut Steamerducks. Flying Steamerduck has been described previously as occurring in our study area only as a non-breeding straggler during the breeding season, or during winter (Humphrey and Livezey 1982, Livezey and Humphrey 1992). Also Chubut Steamerduck is distinct from Flying Steamerduck in a number of characters. In general, the greater bulk, relatively shorter wings, and thicker bill of the Chubut Steamerduck can be distinguished at close range. The more slender neck and longer and more curved tail of Flying Steamerduck can be evident, particularly if the birds adopt the “alert” posture in which the back is arched and neck extended (Humphrey and Thompson 1981, Livezey and Humphrey 1992). When we found pairs and broods, we closely examined the plumage of ducklings and breeding females because these are absolutely diagnostic for identifying each species (Delacour 1975, Livezey and Humphrey 1992, Edelaar and García Borboroglu 2005).
Logistical constraints prevented us from surveying the entire coast. Inclusion or exclusion of survey segments was based on logistical considerations, not *a priori* expectations of the likelihood of finding steamerducks. Surveyed areas were broken into discrete survey segments according to accessibility (cities, towns, landowner permissions).

All surveyed shoreline segments, including island and mainland coasts, were searched using either boat or ground counts, depending on weather conditions and topography. For counts from a boat, we used an inflatable with an outboard motor travelling at a speed of 5–10 knots, between 10 and 15 m from the coast. Ground counts were conducted on foot. With both methods, we used 10 x 40 binoculars and counted any observed steamerducks. Chubut Steamerducks are closely associated with the shoreline, and we assume that we did not miss birds due to their distance from shore. We used GPS to measure the linear distances navigated and walked, both in island and mainland coast segments. The perimeters of the island were estimated using nautical charts (Servicio de Hidrografía Naval Argentina), published literature (García Borboroglu et al. 2002), and GIS. During surveys, steamerducks were identified based on plumage and bill characteristics as adult males and females, breeding pairs, or juveniles. Steamerducks appear to reach sexual maturity at 2–3 years of age (Weller 1972, Delacour 1975, Johnsgard 1978) and the fledging period is about 12 weeks (Weller 1972, Johnsgard 1978). Thus, we consider flocks of non-breeders and post-fledging individuals as juveniles. Previous observations in the last 20 years indicate that, during the breeding season, juvenile groups are consistently located in the same restricted sites (pers. obs.)

To quantify differences between surveys done by boat and from the ground, we used both methods to simultaneously count ducks along 23 coastal segments with varying lengths (0.44 and 4.72 km)

Figure 1. Surveyed coastal sectors are shown as thick lines and distribution range is limited by filled stars. Areas with higher concentrations of individuals: 1. Bahía Melo; 2. Bahía Camarones/Cabo Dos Bahias, and 3. Bahía Bustamante/Caleta Malaspina.
and compared results using the Wilcoxon paired-sample test. The analysis was done for the total number of ducks and for each cohort (adult males and females, breeding pairs, and juveniles). If there were significant differences between survey methods, we developed correction factors based on the ratio of the numbers from boat surveys (which were consistently higher) to the number of ducks from ground surveys. We assumed that boat surveys were as close to a complete census as possible.

**Abundance estimation**

To estimate numbers of steamerducks, we applied correction factors as appropriate based on cohort and survey type (see Results) to survey data, and calculated the number of ducks per kilometre of coast (density) for each surveyed segment. In addition, two different methods were used to extrapolate adult densities to unsurveyed segments within the distribution range of the species. In both, densities of surveyed coastal segments were calculated separately for island segments and mainland segments, and differences were tested using the Mann-Whitney U-test. If differences were statistically significant, island and mainland densities were used separately to extrapolate to the corresponding unsurveyed coastal segments.

**Method # 1.** We calculated an average island density by dividing the total number of adults counted in each island segment by the total number of kilometres of those segments surveyed. Likewise, we used an average to determine mainland density by dividing the number of adults seen by the total number of coastal kilometres surveyed. We used those calculated densities to estimate adult densities and subsequently abundance on unsurveyed island and mainland segments.

**Method #2.** The unsurveyed islands were estimated using the average density from the four closest islands. The abundance of the mainland segments not counted was estimated using an average density calculated from the two adjacent mainland segments surveyed.

**Results**

**Census techniques**

Thirty-nine percent more adults (pairs, males and females combined) were counted from the boat than from land during simultaneous surveys \( z = -2.5, n = 23, P < 0.05; \) Wilcoxon paired-sample test). By boat, we counted 26% more breeding pairs \( z = -2.13, n = 23, P < 0.05; \) Wilcoxon paired-sample test) and 87% more males than when we counted from land \( z = -2.11, n = 23, P < 0.05; \) Wilcoxon paired-sample test). Numbers of females counted from the boat and from land were similar \( z = -0.98, n = 23, P > 0.05; \) Wilcoxon paired-sample test). We used a correction factor for the number of breeding pairs and males when counts were made on land. Numbers of juveniles were similar between boat and ground surveys \( z = -0.45, n = 23, P > 0.05; \) Wilcoxon paired-sample test).

**Breeding distribution**

Chubut Steamerducks have been described previously as a locally common species from Península Valdés \((42°00′S, 65°04′31.11″W)\) south through the Bahía Camarones to the northernmost regions of San Jorge Gulf \((45°14′49.66″S, 66°41′58.43″W)\) along the marine coastlines of Chubut (Humphrey and Thompson 1981, Livezey and Humphrey 1992). However, in our survey, the northernmost and southernmost records of breeding Chubut Steamerducks corresponded to Playa Unión, Rawson \((43°20′38″S, 65°3′19″W)\) and near the Chubut-Santa Cruz Provincial border \((45°59′35″S, 67°35′44″W)\), respectively (Figure 1). Chubut Steamerduck has a restricted and non-uniform breeding distribution in Chubut Province. Steamerducks were more concentrated in the northern San Jorge Gulf sector, which contained about 46% of the adult Chubut Steamerduck population.
The three areas with the highest concentrations of adult ducks were: Bahía Melo (26% of the population, density 7 ducks km\(^{-2}\)); Bahía Camarones/Cabo Dos Bahías (13% of the population, density 8.5 ducks km\(^{-2}\)); and Bahía Bustamante/Caleta Malaspina (9% of the population, density 5.2 ducks km\(^{-2}\); Figure 1).

**Abundance estimation**

We directly observed 1,703 adult Steamerducks along 835 km of surveyed coastline. Applying the corresponding correction factor to segments surveyed from land only, we estimated a total number of adults for surveyed areas to be 1,841 adults. Average (± SE) density of adults around islands (14.6 ± 2.3, range = 0.4–91.6 ducks km\(^{-2}\)) was significantly higher than along the mainland coast (4.1 ± 1.0, range = 0.2–13.6 ducks km\(^{-2}\)) (\(U = 164, P < 0.05\), Mann-Whitney U-test).

We estimated numbers in the unsurveyed area using the weighted average density of islands and mainland coast to be 1,832 adults (Method 1). Using Method 2, we estimated that numbers of adults on unsurveyed segments were 1,587 adults. Therefore, we estimated the total number of adults to be between 3,428 and 3,673 steamerducks.

On the surveys, we recorded 1,899 juveniles, occurring in flocks varying from three to 103 individuals. Juvenile groups were observed in the same sites during several years, some of them reported from at least 1992. We did not mark birds, so it is unknown whether these juvenile groups were necessarily composed of the same individuals across years. Because of their irregular, clumped distribution, we could not extrapolate numbers of juveniles to unsurveyed segments, so this number represents a minimum value.

**Discussion**

Chubut Steamerducks were not uniformly distributed and were concentrated in the northern San Jorge Gulf sector, with a total population likely to be around 3,500 adults and a minimum of approximately 2,000 juveniles.

Chubut Steamerduck densities were higher on islands than mainland shorelines. Steamerducks prefer to nest on islands and islets, mostly along shorelines within bays and inlets with shallow waters protected from waves and predominant winds, and high vegetation cover (Agüero et al. 2010); these characteristics are common features of the San Jorge Gulf, which may explain why Chubut Steamerducks are abundant in this area. Moreover, islands tend to be predator-free, allowing potential higher breeding success for flightless birds (Lack 1968). Several potential mainland mammalian predators have been recorded, which include the Hairy Armadillo *Chaetophractus villosus*, Argentine Grey Fox *Pseudalopex griseus*, Patagonian Ferret *Galictis cuja*, Geoffroy’s Cat *Felis geoffroyi*, and Puma *Puma concolor*. Avian predators are prevalent in island localities. In our study area, two species of birds were observed to predate eggs of Chubut Steamerducks: Kelp Gulls *Larus dominicanus* and Crested Caracara *Polyborus plancus* (Agüero et al. 2010). Great Skuas *Catharacta skua*, Giant Petrels *Macronectes giganteus* and Kelp Gulls prey upon ducklings of *Tachyeres* spp. (Pettingill, 1965). Currently, 38% of the expanding Patagonian population of Kelp Gull (43,875 breeding pairs) breeds in the northern San Jorge Gulf sector, sympatrically with Chubut Steamerducks, on 72% of the islands (Lisnizer et al. 2011).

The recently designated Interjurisdictional Marine Park in San Jorge Gulf contains about 46% of the entire population of Chubut Steamerducks and may provide some protection from disturbance and habitat destruction within its jurisdiction. However, there are other potential threats to Chubut Steamerducks that marine park designation does not protect against. Specifically, oil exploration and extraction are taking place within 100 km of the heart of the population of Chubut Steamerducks. This is a major concern, since Argentina is rapidly developing oil resources on this part of the continental shelf. Moreover, the northward Falkland current would take any spill occurring in the development area directly towards Chubut...
Steamerduck breeding range. In the last 30 years, three major oil spills hit key coastline sectors where steamerducks bred (García Borboroglu et al. 2008), causing massive mortalities of steamerducks (D. Pautaso pers. comm.). No estimations of the number of affected birds were made and, because this study presents the first abundance estimation, there are no population trend data available for the species. The oil spills also harmed the intertidal feeding areas of steamerducks. Contaminants derived from oil are still found in the sediment and on rocky surfaces of both steamerduck breeding and foraging grounds (Gil et al. 1999, Commendatore et al. 2000). These events may undoubtedly have caused a dramatic reduction in the population size and in the quality of habitat, hindering the potential for recovery.

The risk of oil spill due to development or transport can be catastrophic to seabird populations (Bourne 1976, Holmes and Cronshaw 1977, Ohlendorf et al. 1978, Clark 1984). Direct lethal effects of oil contamination and indirect effects of oil spill on marine birds may include: (1) ingestion of, or exposure to, oil that negatively affects reproduction (McGill and Richmond 1979, Lewis and Malecki 1984, Trivelpiece et al. 1984); (2) ingestion of oil that leads to the development of pathological conditions in tissues and reduces overall physical condition (Fry and Lowenstine 1985); (3) loss of food due to prey mortality, and (4) clean-up operations associated with an oil spill (personnel and clean-up methods) that cause a significant disturbance to birds living in shoreline habitats (Maccarone and Bizzorad 1994).

Another current threat is the presence of three introduced invasive species: the Green Crab Carcinus maenus, Asian Kelp Undaria pinnatifida, and the Acorn Barnacle Balanus glandula, all of which are causing dramatic changes to near-shore ecosystems (Casas et al. 2004, Hidalgo et al. 2005, Schwindt 2007) upon which steamerducks rely. The species is also exposed to several other potential anthropogenic threats, including macroalgae and guano harvesting (Piriz and Casas 1996, García Borboroglu and Yorio 2007). Egg-collecting has also been reported but appears to occur at very low intensity at a few sites (P. García Borboroglu unpubl. data). Other human activities that take place along the coast include tourism, recreation and nautical sports (Tagliorette and Losano 1996) and small-scale coastal fishing (Caille 1996).

The new marine park will attract many visitors, enhancing the development of tourist facilities. Therefore it is crucial to design and implement a management plan with a zoning scheme that would protect key breeding areas from harmful activities, such as the construction of piers and ports, the establishment of accommodation facilities, the development of aquaculture farms, the traffic of nautical vessels and terrestrial transport such as ATVs or 4x4 vehicles, and permits to camp, fish or hunt. In addition, the authorities should prepare an effective oil spill contingency plan. Finally, it is imperative to avoid the extraction of adults or eggs from the wild population. Future research should focus on a comprehensive monitoring of the population and on conducting demographic and genetic studies to be able to develop population viability analyses.

Chubut Steamerducks are flightless and have a low number of mature individuals restricted in their distribution and occurring close to the coast. In addition, because they require specific habitat features for breeding and foraging (Gatto et al. 2008, Agüero et al. 2010) they are particularly vulnerable to a wide array of threats. In conclusion, our study, in combination with the current and increased threats mentioned, indicates that the Chubut Steamerduck is of significant conservation concern and suggests that its IUCN listing should be upgraded and listed at least as ‘Vulnerable’.

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